Received: 14 November 2008

(www.interscience.com) DOI 10.1002/jms.1541

# Metabolites from the biodegradation of 1,6-hexanediol dibenzoate, a potential green plasticizer, by *Rhodococcus rhodochrous*

Azadeh Kermanshahi Pour,<sup>a</sup> Orval A. Mamer,<sup>b</sup>\* David G. Cooper,<sup>a</sup> Milan Maric<sup>a</sup> and Jim A. Nicell<sup>c</sup>

Metabolites from the biodegradation of a potential plasticizer 1,6-hexanediol dibenzoate in the presence of *n*-hexadecane as a co-substrate by the common soil organism *Rhodococcus rhodochrous* were identified using GC/MS and Fourier transform mass spectroscopy (FTMS) techniques. Trimethylsilylation of compounds from the biodegradation broth permitted detection of the following metabolites: 1-hexadecyl benzoate, 6-benzoyloxyhexanoic acid, 4-benzoyloxybutanoic acid, 6-benzoyloxyhexan-1-ol and benzoic acid. The presence of these metabolites was confirmed by repeating the biodegradation with 1,6-hexanediol dil<sup>2</sup>H<sub>5</sub>]benzoate, by measurement of their exact masses in FTMS and by comparison with available authentic materials. The results show that biodegradation of 1,6-hexanediol dibenzoate by *R. rhodochrous* does not lead to the accumulation of persistent metabolites as has been reported for commercial dibenzoate plasticizers. Copyright © 2009 John Wiley & Sons, Ltd.

Keywords: 1,6 hexanediol dibenzoate; plasticizer; biodegradation; metabolites; GC/MS; Fourier transform mass spectroscopy

# Introduction

The interaction of microorganisms with xenobiotic chemicals in the environment is a critical issue that must be studied when assessing their toxic impacts on ecological systems and human health.<sup>[1-4]</sup> Incomplete biodegradation of the parent compounds in the environment may lead to their accumulation and the production of metabolites of increased mobility, toxicity and persistence.<sup>[5-8]</sup> Therefore, in order to fully assess the impacts of these xenobiotics in the environment, it is important to identify the full range of compounds that are produced through biodegradation and to assess their toxicity and biodegradability.

In addition, the development of alternative compounds to replace conventional plasticizers requires monitoring of the consequences of their biodegradation to ensure that their metabolites have minimal toxicological impacts in the environment. Plasticizers, which are the most widely used additives in polymer manufacturing,<sup>[9]</sup> have raised serious health and environmental concerns in recent years.<sup>[10,11]</sup> In addition, the concerns raised above about the potential impacts of metabolites have been shown to be the case for a number of widely used commercial plasticizers including phthalates and adipates.<sup>[4,6–8]</sup> Studies of the biodegradation of phthalate and adipate plasticizers have reported the production of several different metabolites with greater toxicity than the parent compounds.<sup>[4,6–8]</sup>

For example, 2-ethylhexanoic acid, a potent peroxisome proliferator,<sup>[12,13]</sup> was identified from the biodegradation of di-2-ethylhexyl adipate, di-2-ethylhexyl phthalate and di-2-ethylhexyl terephthalate.<sup>[6]</sup> Mono-2-ethylhexyl phthalate, a metabolite expected in the biodegradation of di-2-ethylhexyl phthalate,<sup>[14]</sup> is classified as an endocrine disruptor.<sup>[15]</sup> Detection of these and related metabolites of phthalates and adipates in mice, rats and human plasma and urine has led to greater concerns and stricter environmental regulations.<sup>[16–18]</sup>

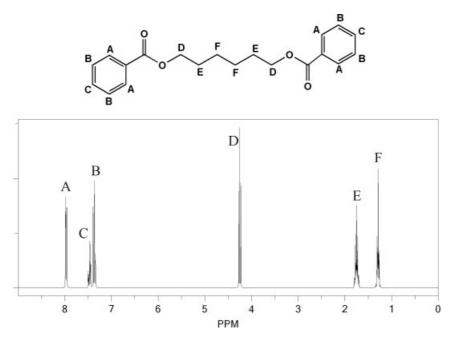
In recent years, dibenzoate plasticizers such as diethylene glycol dibenzoate (D(EG)DB) and dipropylene glycol dibenzoate (D(PG)DB) have been proposed as alternatives to the more commonly used compounds because they tend to degrade more rapidly under the action of common microorganisms.<sup>[19,20]</sup> Although this tendency appears to make them attractive as alternatives to phthalates and adipates, it has been shown that the incomplete microbial hydrolysis of D(EG)DB) and D(PG)DB when microorganisms are growing on glucose as a primary co-substrate leads to the accumulation of diethylene glycol monobenzoate and dipropylene glycol monobenzoate, respectively, which exhibit significant toxicity.<sup>[21]</sup>

However, the rapid degradation of the dibenzoates indicate that a potential route to the development of a 'green' plasticizer may be to start with the basic structure of the more easily degraded dibenzoate plasticizer and modify it to reduce the accumulation of metabolites when undergoing biodegradation. Therefore, the aim of this study was to monitor the biotransformation of 1,6-hexanediol dibenzoate, a potential plasticizer, by *Rhodococcus rhodochrous*, a common soil organism, in the presence of hexadecane as a primary carbon source and to identify all of the metabolites created during biodegradation. Low-resolution

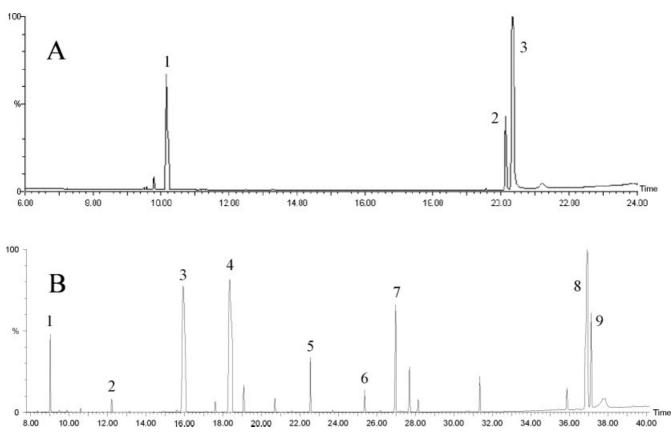
- b Mass Spectrometry Facility, McGill University, 740 Dr Penfield Avenue, Montreal, QC, Canada H3A 1A4
- c Department of Civil Engineering and Applied Mechanics, McGill University, 817 Sherbrooke Street West, Montreal, QC, Canada H3A 2K6

<sup>\*</sup> Correspondence to: Orval A. Mamer, Mass Spectrometry Facility, McGill University, 740 Dr Penfield Avenue, Montreal, QC, Canada H3A 1A4. E-mail: orval.mamer@mcqill.ca

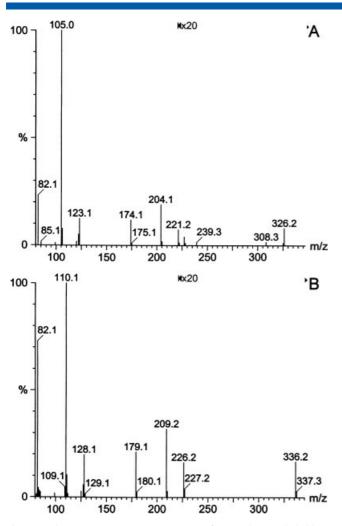
a Department of Chemical Engineering, McGill University, 3610 University Street, Montreal, QC, Canada H3A 2B2



**Figure 1.** NMR spectrum of synthesized 1,6-hexanediol dibenzoate. Relative integration of protons on indicated carbon atom (A:B:C:D:E:F = 2:2:1:2:2:2). A, B, C exhibit the expected patterns for a phenyl group. D and F are triplets and E is a multiplet (assumed to be a triplet of triplets) as expected for this part of the structure.



**Figure 2.** Gas chromatograms obtained for the underivatized broth extract (panel A) and for the extract after trimethylsilylation (panel B). In the former, peaks are identified as 1: hexadecane; 2: 1-hexadecyl benzoate; 3: 1,6-hexanediol dibenzoate. In the latter, peaks are identified as 1: benzoic acid TMS derivative; 2: 1,6-hexanediol TMS derivative; 3: pentadecane added to the extract as a retention time marker; 4: hexadecane added to the broth as a co-metabolite; 5: 4-benzoyloxybutyric acid TMS derivative; 6: 6-benzoyloxyhexan-1-ol TMS derivative; 7: 6-benzoyloxyhexanoic acid TMS derivative; 8: 1,6-hexanediol dibenzoate and 9: 1-hexadecyl benzoate. Other eluting peaks in B originate in the solvent and derivatizing reagent. The elution order for 1-hexadecylbenzoate and 16-hexanediol dibenzoate is reversed by switching between DB-1 and HP-5 columns.



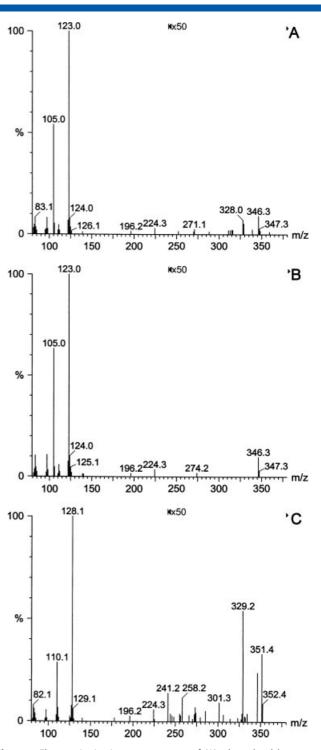
**Figure 3.** Electron ionization mass spectra of (A) 1,6-hexanediol dibenzoate and (B) 1,6-hexanediol di $[{}^{2}H_{5}]$  benzoate isolated from the incubation mixtures. Their mass spectra and GC retention times are identical to those for the diesters synthesized for this study. The intensities of ions above *m/z* 220 are multiplied by 20.

GC/MS with electron ionization was used for the identification of the metabolites as their trimethylsilyl (TMS) derivatives. Fourier transform mass spectroscopy (FTMS) was also used to obtain their underivatized accurate masses with electrospray ionization (ESI).

# **Experimental**

#### **Chemicals and reagents**

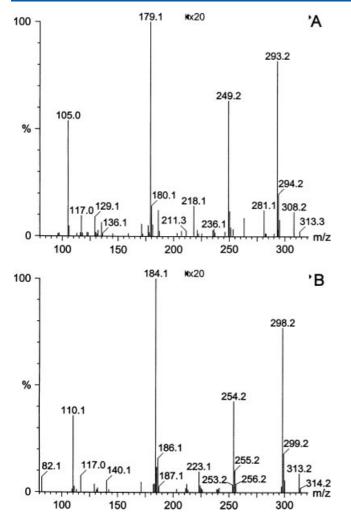
1,6-Hexanediol 99%, 1-hexadecanol 99%, *n*-hexadecane 99% and benzoyl chloride 99% were purchased from Sigma-Aldrich (Oakville, ON, Canada). [ ${}^{2}H_{5}$ ]Benzoyl chloride 99.1 atom% D was purchased from CDN isotopes (Montreal, QC, Canada). Bacto Brain/Heart infusion and yeast extract were obtained from Difco Microbiology (Montreal, QC, Canada) and Fisher Scientific (Montreal, QC, Canada), respectively. Bis(trimethylsilyl)trifluoroacetamide (BSTFA) was purchased from Chromatographic Specialties (Brockville, ON, Canada). Pentadecane was purchased from A & C American Chemicals (Montreal, QC, Canada). All other chemicals were obtained from Fisher Scientific.



**Figure 4.** Electron ionization mass spectra of (A) 1-hexadecyl benzoate, (B) synthesized 1-hexadecyl benzate and (C) 1-hexadecyl  $[^{2}H_{5}]$  benzoate. The intensities of ions above m/z 240 are multiplied by 50.

# Synthesis of 1,6-hexanediol dibenzoate and 1,6-hexanediol di[2H5]benzoate

1,6-Hexanediol dibenzoate was synthesized by refluxing 5 g of 1,6-hexanediol with 20 ml of benzoyl chloride (4 equivalents) under nitrogen in 120 ml of acetone in a round bottom flask for 7 h. The reaction mixture was cooled to room temperature and then diluted with 100 ml of chloroform. The mixture was washed



**Figure 5.** Electron ionization mass spectra of the TMS derivatives of (A) 6-benzoyloxy-hexanoic acid and (B)  $6-[^{2}H_{5}]$  benzoyloxyhexanoic acid isolated from the incubation mixtures. The intensities of ions above m/z 210 are multiplied by 20.

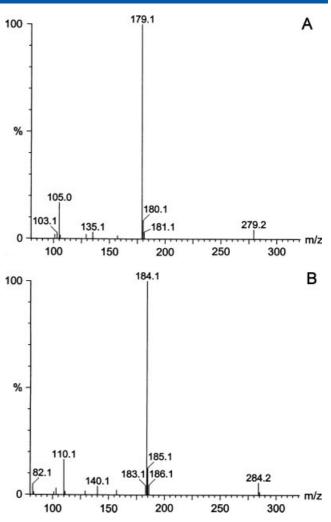
three times with 100 ml of a saturated sodium bicarbonate solution and concentrated to a yellow oil, which, on standing, yielded an off-white powder. This was re-crystallized from heptane.

The synthesis of 1,6-hexanediol di[ ${}^{2}H_{5}$ ]benzoate was achieved in a similar manner by reacting 1,6-hexanediol (0.39 g) with 1.2 ml of benzoyl-[ ${}^{2}H_{5}$ ]-chloride (3 equivalents) in 11 ml of acetone. The procedure used for work-up and re-crystallization is described above.

The proton nuclear magnetic resonance (NMR) spectra of the synthesized 1,6-hexanediol dibenzoate and 1,6-hexanediol di $[{}^{2}H_{5}]$ benzoate were consistent with the spectra expected for these compounds (Fig. 1). The EI spectrum of the unlabelled diester was virtually identical to the spectrum published in the NIST/EPA/NIH 1998 Mass Spectral Library of the United States Department of Commerce.

#### Synthesis of 1-hexadecyl benzoate

1-Hexadecyl benzoate was synthesized by reacting 8 g of 1-hexadecanol with 5 ml of benzoyl chloride (1.3 equivalents) in 120 ml of refluxing acetone in a round bottom flask for 7 h. The reaction was carried out under nitrogen. The reaction



**Figure 6.** Electron ionization mass spectra of the TMS derivatives of (A) 6-benzoyloxyhexan-1-ol and (B)  $6-[^{2}H_{5}]$  benzoyloxyhexan-1-ol isolated from the incubation mixtures.

mixture was cooled to room temperature and then diluted with 100 ml of chloroform. The mixture was washed three times with 100 ml of a saturated sodium bicarbonate solution. The solvent was then evaporated and 1-hexadecyl benzoate was crystallized from the residue. The proton NMR spectrum obtained from the synthesized 1-hexadecyl benzoate was in agreement with the spectrum expected for this compound.

#### **Biodegradation studies**

Biodegradation of 1,6-hexanediol dibenzoate or 1,6-hexanediol di[ ${}^{2}H_{5}$ ]benzoate was conducted in 500 ml Erlenmeyer flasks with a sponge cap. The medium for the experiments consisted of 100 ml of the sterilized minimum mineral salt medium (MMSM) and 0.1 g/l of yeast extract and 2.5 g/l *n*-hexadecane. Either 1,6-hexanediol dibenzoate or 1,6-hexanediol di[ ${}^{2}H_{5}$ ]benzoate (3 mmol/l) were added to the flasks individually before autoclaving. The MMSM contained 4 g/l NH<sub>4</sub>NO<sub>3</sub>, 4 g/l KH<sub>2</sub>PO<sub>4</sub>, 6 g/l Na<sub>2</sub>HPO<sub>4</sub>, 0.2 g/l MgSO<sub>4</sub>·7H<sub>2</sub>O, 0.01 g/l CaCl<sub>2</sub>·2H<sub>2</sub>O, 0.01 g/l FeSO<sub>4</sub>·7H<sub>2</sub>O and 0.014 g/l disodium ethylenediamine-tetraacetic acid.

*Rhodococcus rhodochrous* ATCC 13 808 was obtained from the American Type Culture Collection (ATCC, Rockville, MD, USA) and

was stored at  $-70\,^\circ\text{C}$  in plastic vials containing 20% glycerol and a sterile growth medium of Bacto Brain/Heart infusion broth.

To prepare the initial inoculum, the contents of a vial were thawed and transferred to a 500-ml shaker flask containing sterile growth medium composed of Brain/Heart infusion (30 g/l Brain/Heart infusion broth in 100 ml of distilled water) and then incubated on a rotary shaker (Series 25, New Brunswick Scientific, Edison, NJ, USA) set at 250 rpm and 30 °C. After a day, a new shaker flask containing 100 ml of 30 g/l sterile growth medium of Brain/Heart infusion in distilled water was inoculated with 1 ml of the initial inoculum.

When exponential growth was reached, this microbial culture was used to inoculate 100 ml of the sterilized MMSM containing 0.1 g/l yeast extract and 2.5 g/l *n*-hexadecane. This was used to inoculate the shaker flasks containing either 1,6-hexanediol dibenzoate or 1,6-hexanediol di[ $^{2}H_{5}$ ]benzoate for the biodegradation study. The shaker flasks were incubated for a period of 7 days on a rotary incubator shaker set at 250 rpm and 30 °C.

#### Sample preparation for GC/MS and GC/FID analyses

Over the course of biodegradation of 1,6-hexanediol dibenzoate, triplicate samples of 3 ml each were taken from the biodegradation broth every day. The samples were adjusted to pH 2 through the addition of sulfuric acid and extracted with 3 ml of chloroform. For GC/MS analysis, the extracts were evaporated to dryness under a dry nitrogen stream and the residues were taken up in 50  $\mu$ l of anhydrous pyridine. Trimethylsilyl (TMS) derivatives were prepared by the addition of 50  $\mu$ l of BSTFA to the pyridine solutions in capped auto-injector vials, which were heated in an aluminum block at 60 °C for 15 min. For GC/ flame ionization detection (FID) analysis, chloroform extracts of the samples were used without derivatization.

#### GC/MS analyses

Aliquots (1  $\mu$ I) of the underivatized extracts were analyzed in lowresolution GC/MS mode with a GCT (Micromass, Manchester, UK) fitted with a 30-m HP-5 capillary column having a 0.32-mm i.d. and 0.25- $\mu$ m film thickness. The temperature was programmed from 80 °C after 1-min hold to 300 °C at 10 °C/min followed by a bake-out period of 6 min at 300 °C. The injector was operated in 1:100 split mode at 250 °C with a constant helium pressure of 70 kPa. The GC re-entrant temperature was 250 °C. The EI ion source was operated at 70 eV and 200 °C.

TMS-derivatized extracts and the synthesized 1-hexadecyl benzoate were analyzed in GC/MS mode on a 30-m, 0.25-mm i.d. DB-1 column operated as described above. The scan range was m/z 80–600 to avoid the intense but uninformative m/z 73 common to TMS derivatives.

## **FTMS** analyses

High-resolution measurements of the underivatized extracts were made in positive ion electrospray mode with an IonSpec 7.0 tesla FTMS (Lake Forest, CA, USA) calibrated with polyethylene glycol 300. The instrument was equipped with a 'Z'-spray source from Waters Corporation (Milford, MA, USA), an accumulation hexapole, a collision cell, a hexapole ion guide, a standard cylindrical ion cyclotron resonance (ICR) cell and Omega 9 software. The analyses used a direct infusion flow rate of  $2-3 \mu$ l/min in solution with 90:10 v/v methanol: water. Formic acid (1%) and sodium iodide were

added to enhance cationization and to provide a secondary mass scale calibrating ion (Na<sub>2</sub>I<sup>+</sup>, *m/z* 172.8835). The 'Z'-spray source used capillary and cone voltages of 3899 and 30 V, respectively. lons were accumulated in the hexapole for 300–1500 ms with a rod voltage of 70 V. For the transfer of ions to the ICR cell through the hexapole ion guide, the low mass range coil with a frequency of 3020 kHz was used along with a voltage of 80 V. For detection, ions were excited through an arbitrary waveform in a range of *m*/*z* 100–1000 with an amplitude of 135 V(b-p); the analog to digital conversion (ADC) rate for the MS was 2 MHz for a scan range of *m*/*z* 75–500. Transients were 1M data points long. A waiting time of 5 s before the detection step was used to allow the pressure in the ICR cell to return to its nominal value of 2 × 10<sup>-9</sup> torr.

#### GC/FID analyses

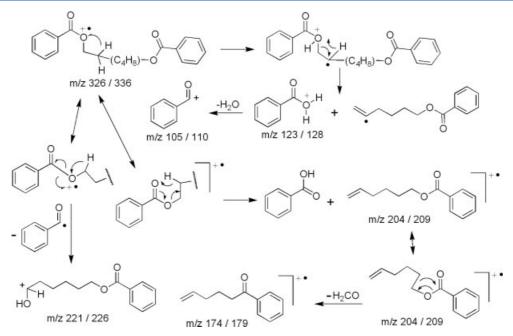
Aliquots (1  $\mu$ l) of the chloroform extracts were analyzed in a Varian CP-3800 gas chromatograph equipped with a 30 m  $\times$  0.32 mm i.d. fused silica 8CB column (Varian, Montreal, QC, Canada) programmed to 300 °C at 10 °C after a 2-min hold initially at 40 °C. The injection port and FID were kept at 250 °C and 300 °C, respectively. Helium was used as a carrier gas at a flow rate of 1.5 ml/min. The concentration of the metabolites were estimated by GC/FID.

# **Results and Discussion**

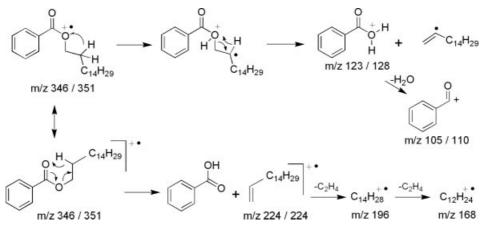
1,6-Hexanediol dibenzoate was synthesized in an attempt to develop a more environmentally benign version of the standard dibenzoate plasticizers. It has been shown that analogous compounds can be converted to stable toxic metabolites by microorganisms while growing on an easily metabolized substrate.<sup>[21]</sup> The purpose of this work was to determine whether co-metabolism of 1,6-hexanediol dibenzoate by a typical soil microorganism *R. rhodochrous* resulted in the production of stable metabolites.

Figure 2 shows the total ion current GC of an extract of an experiment using 1,6-hexanediol dibenzoate without derivatization (panel A) and after trimethylsilylation (panel B). Retention times and elution order are different in panel A and panel B due to the use of a HP-5 column and 10 °C program (panel A) *versus* a DB-1 column and a 5 °C temperature program for the derivatized sample (panel B).

Figure 3A and B shows the El mass spectra of 1,6-hexanediol dibenzoate and 1,6-hexanediol di[<sup>2</sup>H<sub>5</sub>]benzoate, respectively. Weak molecular cations at m/z 326 and 336 were observed in their mass spectra, respectively. Fragmentation pathways are proposed in Scheme 1. Fragment ions at m/z 204/209 are formed by loss of benzoic acid via a McLafferty mechanism and with a possible subsequent elimination of the elements of formaldehyde to yield ions at m/z 174 and 179. Formaldehyde elimination, as proposed in Scheme 1, is not observed in the dibenzoates of 1,5-pentanediol or 1,4-butanediol (data not shown, available in the NIST/EPA/NIH 1998 Mass Spectral Library), and this may be related to the alkyl chain length. There are three possible precursor ions to m/z174/179; these are *m*/*z* 221/226, 204/209 and the M<sup>+•</sup>. The first would violate the even electron 'rule', and it is not possible to decide between the last two on the basis of the present data. Fragments at m/z 123/128, nominally protonated benzoic acid, may be formed in a four-centred elimination of a radical olefin. Alternatively, m/z 123/128 may have the Ph-C(OH)<sub>2</sub><sup>+</sup> structure. Subsequent loss of water by m/z 123/128 results in ions at m/z



Scheme 1. Proposed fragmentation scheme for 1,6-hexanediol dibenzoate and 1,6-hexanediol di $[{}^{2}H_{5}]$  benzoate. The second m/z value refers to the labelled diester.



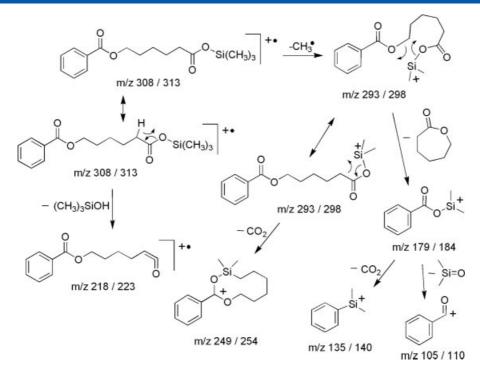
Scheme 2. Proposed mass spectrometric fragmentation scheme for 1-hexadecyl benzoate and 1-hexadecyl [<sup>2</sup>H<sub>5</sub>]benzoate. The second *m/z* value refers to the labelled ester.

105/110. Formation of this last ion by a direct C(O)–O fission in the molecular ion cannot be discounted.

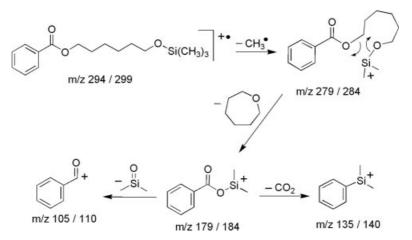
Figure 4A and C is the mass spectra of 1-hexadecyl benzoate and 1-hexadecyl [ ${}^{2}H_{5}$ ]benzoate, respectively, isolated from the broth. The mass spectrum of the synthesized 1-hexadecyl benzoate is shown in Fig. 3B. The proposed fragmentation scheme for 1-hexadecyl benzoate and 1-hexadecyl [ ${}^{2}H_{5}$ ]benzoate in Scheme 2 accounts for the formation of fragments at m/z 123/128 and 105/110 in a manner parallel to that suggested in Scheme 1 for 1,6-hexanediol dibenzoate and 1,6-hexanediol di[ ${}^{2}H_{5}$ ]benzoate. Again, as in Scheme 1, although m/z 123/128 is drawn as protonated benzoic acid, the ion structure may be that of a  $\alpha, \alpha$ -dihydroxybenzyl cation. m/z 105/110 may be formed by H<sub>2</sub>O loss or directly from the M<sup>+•</sup>. A McLafferty rearrangement in the molecular cations generates m/z 224/224, which undergo the sequential loss of two ethylene groups.

Figure 5A and B shows the mass spectra of the trimethylsilylated metabolites identified as 6-benzoyloxyhexanoic acid and 6-[<sup>2</sup>H<sub>5</sub>]benzoyloxyhexanoic acid, respectively. In support of this assignment, Scheme 3 accounts for the major ion fragments and deuterium labelling found in Fig. 5A and B. Loss of methyl radical by the molecular radical cations yields *m/z* 293 and 298. Subsequent loss of CO<sub>2</sub> by *m/z* 293/298 yields *m/z* 249/254. *M/z* 179/184 is formed by the loss of the elements of  $\varepsilon$ -caprolactone by *m/z* 293/298, and goes on to lose CO<sub>2</sub> to form *m/z* 135/140 and Si(CH<sub>3</sub>)<sub>2</sub>O resulting in *m/z* 105/110. *M/z* 105/110 formation directly from the molecular cation can neither be excluded here nor in the case of Scheme 4. *M/z* 117 is an ion characteristic of TMS derivatives of compounds with carboxy groups, which corresponds to <sup>+</sup>COOTMS.<sup>[22]</sup> Weak *m/z* 117 ions were observed in both Fig. 5A and B.

Figure 6A and B represents the spectra of the TMS derivatives of 6-benzoyloxyhexan-1-ol and  $6-[^{2}H_{5}]$ benzoyloxyhexan-1-ol, which were expected metabolites in the biodegradation of 1,6-hexanediol dibenzoate and 1,6-hexanediol di $[^{2}H_{5}]$ benzoate, respectively, by analogy to the metabolites reported for related



**Scheme 3.** Proposed mass spectrometric fragmentation scheme for the TMS derivatives of 6-benzoyloxyhexanoic and  $6-[^{2}H_{5}]$  benzoyloxyhexanoic acids. The second m/z value refers to the labelled TMS ester.



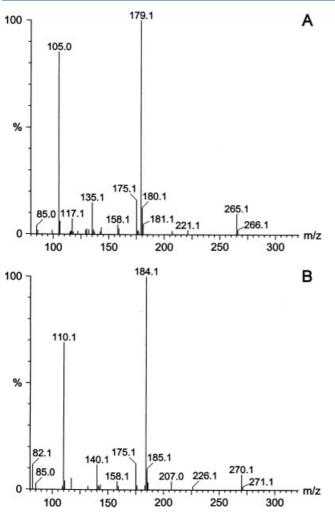
Scheme 4. Proposed mass spectrometric fragmentation scheme for the TMS derivatives of 6-benzoyloxyhexan-1-ol. The second *m/z* value refers to the labelled TMS ester.

commercial plasticizers.<sup>[21]</sup> Scheme 4 proposes fragmentations that account for the major ions in the mass spectra. An interesting Me<sub>2</sub>Si migration appears to occur in the loss of CO<sub>2</sub> for the transition m/z 179/184–135/140. The latter ion is drawn as a silicon analogue of an  $\alpha$ , $\alpha$ -dimethylbenzyl cation, but the actual structure is not known. M/z 135 is also an intense fragment in the spectrum of the TMS ester of benzoic acid (NIST/EPA/NIH 1998 Mass Spectral Library) and is similarly formed by loss of CO<sub>2</sub> from m/z 179 produced by methyl radical loss from the molecular cation (m/z 194), both metastable confirmed (data not shown).

The spectra of the TMS derivatives of 4-benzoyloxybutanoic acid and 4-[ ${}^{2}H_{5}$ ] benzoyloxybutanoic acids are illustrated in Fig. 7A and B, and show homology with the spectra of 6-benzoyloxyhexanoic and 6-[ ${}^{2}H_{5}$ ]benzoyloxyhexanoic acids (Fig. 5A and B). *M/z* 265 and 270 are formed by the loss of methyl radical by the molecular radical cations (m/z 280 and 285, not detected), and go on to lose CO<sub>2</sub> forming m/z 221 and 226. lons at m/z 179/184, 135/140 and 105/110 likely have the same structures as those proposed in Scheme 3. M/z 117 was also observed in both Fig. 7A and B indicating the presence of a carboxyl group.

Figure 8A and B is the mass spectra of the TMS derivatives of benzoic and  $[{}^{2}H_{5}]$ benzoic acids, respectively, found in the derivatized extracts. The molecular ions at m/z 194 and 199 show fragmentations similar to those found in Figs 5, 6 and 7.

The experimental and calculated exact masses obtained for 1,6-hexanediol dibenzoate and 1,6-hexanediol di $[{}^{2}H_{5}]$ benzoate and their metabolites are presented in Table 1. The difference between the calculated and experimental masses was 2.3 ppm or better.

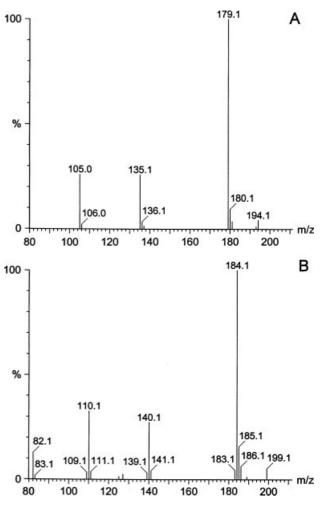


**Figure 7.** Electron ionization mass spectra of the TMS derivatives of (A) 4-benzoyloxybutanoic acid and (B) 4-[<sup>2</sup>H<sub>5</sub>] benzoyloxybutanoic acid isolated from the incubation mixtures.

None of the above metabolites were observed in abiotic control experiments, eliminating the possibility of the formation of any of these metabolites by chemical hydrolysis or oxidation.

Table 2 contains data for the highest observed concentrations and maximum lifetime for each of the metabolites. All of these metabolites eventually disappeared. The most important of these is the monoester, 6-benzoyloxyhexan-1-ol. This compound is analogous to the monoesters produced by co-metabolism of the commercial plasticizers D(EG)DB and D(PG)DB.<sup>[21]</sup> However, the monoesters (diethylene glycol monobenzoate and dipropylene glycol monobenzoate) from biodegradation of the commercial plasticizers were not only resistant to further biodegradation but were also shown to exhibit significant toxicity in screening assays.<sup>[21]</sup> Therefore, the fact that the monoester of 1,6-hexanediol was only observed in small quantities and also degraded rapidly supports the hypothesis that this compound may represent a more environmentally benign dibenzoate plasticizer.

The most long lived of the metabolites in Table 2 and one that was observed at an order of magnitude greater concentration than the monoester was 1-hexadecyl benzoate. This is the only metabolite that cannot originate directly from the degradation of 1,6-hexanediol dibenzoate. It is hypothesized that the formation of 1-hexadecyl benzoate is the result of an enzymatic conjugation



**Figure 8.** Electron ionization mass spectra of the TMS derivatives of (A) benzoic acid and (B)  $[^{2}H_{5}]$  benzoic acid isolated from the incubation mixtures.

of benzoate hydrolyzed from 1,6-hexanediol dibenzoate or its metabolites to hexadecanol, a metabolite of hexadecane degradation. This is consistent with the fact that biodegradation of 1,6-hexanediol di[ ${}^{2}H_{5}$ ]benzoate resulted in formation of 1-hexadecyl [ ${}^{2}H_{5}$ ]benzoate.

This ester of 1-hexadecanol is not a potential problem. It is an artefact arising from the use of hexadecane as the primary carbon and energy source. Hexadecane was convenient for these experiments because *R. rhodochrous* grows well on hydrocarbons<sup>[4]</sup> and a hydrophobic substrate helps to disperse the water-insoluble plasticizer. However, in an environmental situation, this benzoate is unlikely to be formed because there will not be appreciable amounts of alkanes or alcohols present.

# Conclusions

GC/MS and FTMS were used to identify metabolites arising from the biodegradation of 1,6-hexanediol dibenzoate by *R. rhodochrous*. All of these metabolites were confirmed by repeating the experiments with deuterium-labelled analogues.

In contrast to commercially available dibenzoate plasticizers, metabolism of 1,6-hexanediol dibenzoate did not result in accumulation of persistent metabolites. Furthermore, the most

Compound	Ion Composition	Found	Required	Error (ppm)
1,6-Hexanediol dibenzoate	$C_{20}H_{22}O_4Na$	349.1410	349.1410	0.0
	C <sub>20</sub> H <sub>23</sub> O <sub>4</sub>	327.1591	327.1591	0.0
1,6-Hexanediol di[ <sup>2</sup> H <sub>10</sub> ]benzoate	$C_{20}H_{12}{}^2H_{10}O_4Na$	359.2037	359.2038	0.3
	C <sub>20</sub> H <sub>13</sub> <sup>2</sup> H <sub>10</sub> O <sub>4</sub>	337.2222	337.2218	1.1
1-Hexadecyl benzoate	C <sub>23</sub> H <sub>39</sub> O <sub>2</sub>	347.2953	347.2945	2.3
1-Hexadecyl [ <sup>2</sup> H <sub>5</sub> ]benzoate	C <sub>23</sub> H <sub>34</sub> <sup>2</sup> H <sub>5</sub> O <sub>2</sub>	352.3252	352.3258	1.7
6-Benzoyloxyhexanoic acid	C <sub>13</sub> H <sub>16</sub> O <sub>4</sub> Na	259.0942	259.0941	0.4
6-[ <sup>2</sup> H <sub>5</sub> ]Benzoyloxyhexanoic acid	C <sub>13</sub> H <sub>11</sub> <sup>2</sup> H <sub>5</sub> O <sub>4</sub> Na	264.1254	264.1255	0.4
6-Benzoyloxyhexan-1-ol	C <sub>13</sub> H <sub>18</sub> O <sub>3</sub> Na	245.1148	245.1148	0.0
6-[ <sup>2</sup> H <sub>5</sub> ]Benzoyloxyhexan-1-ol	$C_{13}H_{13}{}^{2}H_{5}O_{3}Na$	250.1457	250.1462	2.0
4-Benzoyloxybutanoic acid	C <sub>11</sub> H <sub>12</sub> O <sub>4</sub> Na	231.0628	231.0628	0.0
4-[ <sup>2</sup> H <sub>5</sub> ]Benzoyloxybutanoic acid	$C_{11}H_7^2H_5O_4Na$	236.0942	236.0942	0.0
Benzoic acid	C <sub>7</sub> H <sub>7</sub> O <sub>2</sub>	123.0441	123.0441	0.0
[ <sup>2</sup> H₅]Benzoic acid	$C_7 H_2^2 H_5 O_2$	128.0757	128.0754	2.3

Table 2.	Metabolites	from	the	biodegradation	of	1,6-hexanediol
dibenzoate by R. rhodochrous <sup>a</sup>						

,								
Metabolites	Highest concentration observed, mmol/l <sup>b</sup>	Time of observation, h	Time of confirmed disappearance, h					
6-Benzoyloxyhexan-1- ol	$\textbf{0.03} \pm \textbf{0.01}$	40	87					
6-Benzoyloxyhexanoic acid	$\textbf{0.30}\pm\textbf{0.04}$	40	87					
4-Benzoyloxybutanoic acid	$\textbf{0.07} \pm \textbf{0.01}$	64	87					
Benzoic acid	$\textbf{0.20} \pm \textbf{0.02}$	40	87					
1-Hexadecyl benzoate	$0.31\pm0.01$	40	185					

<sup>a</sup> Initial concentration of 1,6-hexanediol dibenzoate was 3 mmol/l.

<sup>b</sup> Values are the average of triplicate samples. Samples were taken once per day.

stable of the metabolites would not be expected to be observed in the environment. These results support the potential to use 1,6-hexanediol dibenozate as a green plasticizer.

#### Acknowledgements

The authors are grateful for the financial support of the Natural Sciences and Engineering Research Council of Canada and the EJLB Foundation of Canada. We would also like to thank Mr Luc Choinière, Dr Alain Lesimple and Mr Ranjan Roy for GC/MS and FTMS analysis. O.M. wishes to acknowledge grants from the Canadian Institutes for Health Research and the Canadian Foundation for Innovation. Azadeh Kermanshahi Pour thanks NSERC, the McGill Engineering Doctoral Award program, and the Eugenie Ulmer Lamothe fund of the Department of Chemical Engineering at McGill University for providing scholarships in support of her Ph.D. studies.

## References

[1] D. T. Gibson. Microbial degradation of aromatic compounds. Science 1968, 161, 1093.

- [2] W. J. G. M. Peijnenburg. Structure-activity relationship for biodegradation: a critical review. Pure and Applied Chemistry 1994, 66.1931.
- [3] J. W. Raymond, T. N. Rogers, D. R. Shonnard, A. A. Kline. A review of structure-based biodegradation estimation methods. Journal of Hazardous Material 2001, 84, 189.
- [4] A. Nalli, D. G. Cooper, J. A. Nicell. Interaction of metabolites with R. rhodochrous during the biodegradation of di-ester plasticizers. Chemosphere 2006, 65, 1510.
- [5] G. Zink, K. E. Lorber. Mass spectral identification of metabolites formed by microbial degradation of polycyclic aromatic hydrocarbons (PAH). Chemosphere 1995, 31, 4077.
- [6] S. Nalli, D. G. Cooper, J. A. Nicell. Biodegradation of plasticizers by Rhodococcus rhodochrous. Biodegradation 2003, 13, 343.
- [7] O. Horn, S. Nalli, D. G. Cooper, J. A. Nicell. Plasticizer metabolites in the environment. Water Research 2004, 38, 3693.
- [8] A. Nalli, D. G. Cooper, J. A. Nicell. Metabolites from the biodegradation of di-ester plasticizers by Rhodococcus rhodochrous. Science of the Total Environment 2006, 366, 286.
- [9] G. Wypych. Plasticizer types. In The Handbook of Plasticizers, G. Wypych (ed). ChemTec Laboratories Inc.: Toronto, 2004, 7.
- [10] C. S. Giam, E. Atlas, M. A. Powers Jr, J. E. Leonard. Phthalic acid esters. In The Handbook of Environmental Chemistry: Anthropogenic Substances, vol. 3, Hutzinger, O. (ed). Springer-Verlag: Berlin, 1984, 67.
- [11] W. R. Roy. The environmental fate of plasticizer. In *The Handbook* of Plasticizers, G. Wypych (ed). ChemTec Laboratories Inc.: Toronto, 2004.591.
- [12] M. C. Cornu, J. C. Lhuguenot, A. M. Brady, R. Moore, M. C. Elcombe. Identification of the proximate peroxisome proliferator(s) derived from di(2-ethylhexyl) adipates and species differences in response. Biochemical Pharmacology 1992, 43, 2129.
- [13] Y. Keith, M. C. Cornu, P. M. Canning, J. Foster, J. C. Lhuguenot, C. R. Elcombe. Peroxisome proliferation due to di (2-ethylhexyl) adipate, 2-ethylhexanol and 2-ethylhexanoic acid. Archives of Toxicology 1999, 66, 321.
- [14] C. A. Staples, D. R. Peterson, T. F. Parkerton, W. J. Adams. The environmental fate of phthalate esters: a literature review. Chemosphere 1997, 35, 667.
- [15] T. M. Onorato, P. W. Brown, P. Morris. Mono-(2-ethylhexyl)phthalate increase spermatocyte mitochondrial peroxiredoxin 3 and cyclooxygenase 2. Journal of Andrology 2008, 29, 293.
- [16] A. M. Mitchell, J. C. Lhuguenot, J. W. Bridges, C. R. Elcombe. Identification of the proximate peroxisome proliferator(s) derived from di(2-ethylhexyl) phthalate. Toxicology and Applied Pharmacology 1985, 80, 23.
- [17] H. G. Wahl, Q. Hong, S. Hildenbrand, T. Risler, D. Luft, H. Liebich. 4-Heptanone is a metabolite of the plasticizer di(2-ethylhexyl) phthalate (DEHP) in haemodialysis patients. Nephrology dialysis transplantation 2004, 19, 2576.



- [18] S. Sathyanarayana, C. J. Karr, P. Lozano, E. Brown, A. M. Calafat, F. Liu, S. H. Swan. Baby care products: possible sources of infant phthalate exposure. *Pediatrics* **2008**, *121*, 260.
- [19] J. Lang, B. E. Stanhope. Benzoate plasticizer for flexible PVC injection molded toy applications. *Plastics Additives and Compounding* **2001**, 3, 30.
- [20] X. Peng, X. Zhang. Application of a new plasticizer diethylene glycol dibenzoate (DEDB) in poly (vinyl chloride) products. *Zhongguo Suliao* 2000, 14, 64.
- [21] J. Gartshore, D. G. Cooper, J. A. Nicell. Biodegradation of plasticizers by *Rhodotorula Rubra*. *Environmental Toxicology and Chemistry* **2002**, *22*, 1244.
- [22] L. K. Ng, M. Hupé. Simple gas chromatographic method for the assay of salts of carboxylic acids as their trimethylsilyl derivatives. *Journal of Chromatography* **1993**, *637*, 104.