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Journal of the Air & Waste Management Association

Publication details, including instructions for authors and subscription information: http://www.tandfonline.com/loi/uawm20

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To cite this article: Christian J. Murray & Charles R. Nelson (2000) State-Space Modeling of the Relationship between Air Quality and Mortality, Journal of the Air & Waste Management Association, 50:7, 1075-1080, DOI: 10.1080/10473289.2000.10464158

To link to this article: http://dx.doi.org/10.1080/10473289.2000.10464158

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State-Space Modeling of the Relationship between Air Quality and Mortality

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ABSTRACT

A portion of a population is assumed to be at risk, with the mortality hazard varying with atmospheric conditions including total suspended particulates (TSP). This at-risk population is not observed and the hazard function is unknown; we wish to estimate these from mortality count and atmospheric variables. Consideration of population dynamics leads to a state-space representation, allowing the Kalman Filter (KF) to be used for estimation. A harvesting effect is thus implied; high mortality is followed by lower mortality until the population is replenished by new arrivals.

The model is applied to daily data for Philadelphia, PA, 1973–1990. The estimated hazard function rises with the level of TSP and at extremes of temperature and also reflects a positive interaction between TSP and temperature. The estimated at-risk population averages about 480 and varies seasonally. We find that lags of TSP are statistically significant, but the presence of negative coefficients suggests their role may be partially statistical rather than biological. In the population dynamics framework, the natural metric for health damage from air pollution is its impact on life expectancy. The range of hazard rates over the sample period is 0.07 to 0.085, corresponding to life expectancies of 14.3 and 11.8 days, respectively.

IMPLICATIONS

Mortality in an at-risk population may be modeled in a state-space framework. The relation of the hazard rate to TSP and other variables and the unobserved population are estimated using the KF and maximum likelihood. The results for daily data from Philadelphia suggest that both TSP and temperature are risk factors, the effect of both rising at high temperature levels. The appropriate metric for health damage is the effect of particulates on life expectancy, which in the context of daily variation is estimated to be on the order of days.

INTRODUCTION

An extensive literature documents evidence that episodes of severe air pollution are associated with heightened mortality in urban populations. However, the question of whether mortality occurs within a relatively limited population of frail individuals, or whether the risks are more diffuse across the general population, has a long history and is still being examined and debated by researchers. ¹⁻³

In this paper, we explore a new approach to modeling the relationship between mortality and air quality and apply it to daily data from Philadelphia, PA. We take as given that a portion of the urban population is at risk, subject to a probability of death, or hazard rate, that varies with atmospheric conditions including total suspended particulates (TSP). New entrants replenish the at-risk population over time. Only mortality and atmospheric conditions are observed; the at-risk population, new entrants, and the hazard rate are unobserved. Because of the dependence of mortality on the unobserved at-risk population, the model is not directly estimable as a regression, but fortunately it may be cast in state-space form. Using the Kalman Filter (KF) we then infer the hazard rate over time, its relationship to atmospheric variables, and the implied path of the unobserved at-risk population. In the state-space framework, one is able to address the following questions: What is the size of the at-risk population? What is the life expectancy of individuals in that population? What is the effect of changes in air quality on that life expectancy?

The implications of the at-risk population model in state-space form differ fundamentally from those of linear models based on multiple regression.⁴ In the latter, a risk factor such as TSP affects mortality directly, and a persistent rise in the level of TSP would imply a persistent rise in mortality count. In the state-space model the impact of a risk factor such as TSP on mortality is indirect; it is proportional to the size of the at-risk population on

that day which is not observed directly. If the at-risk population has been depleted by recent mortality, then the impact of risk factors will be mitigated since it is a temporarily smaller population that is at risk. Indeed, it is this "harvesting effect" which allows the unobserved at-risk population to be estimated by the KF. If the higher hazard rate persists, the mortality count will fall back toward its previous level, since in the long run mortality is limited to the rate of new arrivals. However, the life expectancy of those in the at-risk population will fall, and this paper presents estimates of that effect.

MODEL SPECIFICATION AND ESTIMATION

Any population is diminished by deaths and replenished by the arrival of new entrants; in this case, we studied the population of individuals in Philadelphia who are at risk from changing air quality. Using P_t to denote the number of individuals in the at-risk population at the end of day t, D_t the number of deaths (mortality) during day t, and N_t the number of new entrants to the at-risk population during day t, the basic equation of population dynamics is

$$P_t = P_{t-1} + N_t - D_t (1)$$

Equation 1 may be viewed as simply an accounting relation that states that the at-risk population at time *t* is equal to its previous value plus the number of new entrants less the number of deaths. The only variable in this relationship that is actually observed in the present context is the mortality count.

Mortality is influenced by atmospheric variables through a hazard function that operates on the at-risk population. Listing atmospheric variables in a vector denoted x_t , we assume the hazard function to be the linear combination of these variables, denoted $\gamma' x_{t}$. The atmospheric variables listed in x_{t} will include measures of air pollution such as TSP and weather variables such as average temperature (AVGT). The elements of the vector γ are coefficients that indicate how each atmospheric variable affects the hazard function. These coefficients are unknown and will have to be estimated from data. The hazard rate is the value of the hazard function at period t, and is the expected fraction of deaths in the at-risk population on that date. Some deaths will result from other factors that affect mortality but which we have not included in x_{t} , so there will also be an error term, e_{t} , which is the difference between actual and expected mortality.

We summarize this discussion by the expression

$$D_t = (\gamma' x_t) \bullet P_{t-1} + e_t \tag{2}$$

A constant in the hazard function captures the average level of mortality, so the mean value of the error term is zero. We seek to estimate the γ vector of coefficients and the time series of the unobserved at-risk population P from daily data on mortality and atmospheric variables.

Note that eq 2 implies that mortality is a nonlinear function of P_{t-1} and the atmospheric variables, the expected number of deaths being the product of hazard and population. This implies that mortality cannot be expressed as a linear function of current and past atmospheric variables with constant coefficients. Indeed, the response of mortality to a change in atmospheric conditions depends on the time path of past conditions, reflected in the current level of the at-risk population.

To complete the model we need to say how N_{t} , the unobserved number of new entrants to the at-risk population, evolves through time. In any population, it is the rate of arrival of new entrants that determines mortality over a long time span. The time series of mortality of those over 65 in Philadelphia displays remarkably little longterm variation over the 14-yr period of our sample. The annual averages fluctuate in a narrow range around the long-run average of 35, ranging from 33 to 37, and ending the period at the upper end of that range. Further, Dickey-Fuller tests strongly reject non-stationarity or drift over long time horizons. Given the long-run relationship between arrivals and mortality, these conditions all suggest that the rate of arrivals fluctuates around a long-run mean during this period in the Philadelphia population. With this background, we employ the specification

$$N_t = N + \eta_t \tag{3}$$

where N is the average number of arrivals, and η_t is a random deviation.

The fundamental problem of statistical inference in this setting is that the at-risk population is an unobserved variable. If $P_{t,1}$ were observed, then one could estimate eq 2 by least squares, regressing mortality on variables that would be cross-products of the x's and $P_{t,i}$. However, $P_{t,l}$ is not observed, so the model is non-linear and cannot be expressed as a linear regression. Fortunately, this model is a member of the class of state-space models that can be made operational using the KF. Briefly, the statespace representation consists of two equations, a measurement equation and a state equation. The former shows how the variable we observe and wish to explain depends on unobserved variables called state variables. The latter shows how those state variables evolve through time. Thus, eq 2 is the measurement equation and eqs 1 and 3 are combined to form the state equation.

The random shocks in these two equations are assumed to be uncorrelated in the state-space formulation. Those shocks are errors in predicting mortality and random arrivals to the at-risk population, respectively. The

assumption that they are uncorrelated is equivalent to saying that there is a well-defined at-risk population in the sense that the movement of individuals from the population at large to the at-risk population does not depend on mortality. We want the model to estimate the size of the at-risk population, so we would not want to exclude a portion of the population that is also at risk. Given parameter values, we obtain an estimate of the series P_t by "running" the data through the KF. 5.6 The resulting estimated level of P_t denoted $P_{t|t'}$ is the minimum mean squared estimate of P_t based on data up to time t and is referred to as the filtered estimate.

A byproduct of the KF is the construction of the Gaussian likelihood function using Harvey's predictionerror decomposition.^{7,8} Maximum likelihood estimation is carried out using the OPTMUM module of the GAUSS programming language. The prior distribution for the initial value of the unobserved at-risk population was given a mean of 0 and variance of 1000, a very loose prior distribution that leaves determination of the at-risk population to be determined by the data. No prior distribution is required for N_t since it is just a constant plus a random error. Maximum likelihood estimates of the parameters are computed by maximizing the likelihood function over the parameter space. This requires that we assume that the error terms in eqs 2 and 3 are normal, serially random, and uncorrelated with each other as well as with variables in the model. The error terms cannot be exactly normal in this context since mortality is integer-valued and non-negative, so what we have is an approximation to maximum likelihood.

Standard errors for parameter estimates are based on asymptotic distribution theory that may not be exact in finite samples, but sample size here is very large. The parameters to be estimated are the γ vector, the average arrival rate N, and the standard deviations of the two error terms e and η denoted σ_e and σ_n , respectively. The t-statistic shown in parentheses beside each point estimate in the tables is based on numerical second derivatives to obtain standard errors. The value of the log of the likelihood function is also reported, since it gives us a basis for comparing models using likelihood ratio tests or information criteria. For nested models, the likelihood ratio test is that 2 times the difference in log likelihood will be approximately chi-square with degrees of freedom equal to the number of additional parameters in the more complex model.

We have verified that the algorithm does in fact uncover the underlying parameters and the at-risk population in a system with the structure of eqs 1–3 by confronting it with artificial mortality data that we generated. Our hazard function was $(0.1 + 0.003 \cdot x_t)$ where the sequence $\{x_t\}$ was constructed to mimic the behavior

of TSP in the actual data. N was set at the mean mortality rate in the real data, and the standard deviations of the error terms were set at 2 and 3, respectively. Series length was 6000 to be comparable to the daily data we studied. To increase the validity of the experiment, one of us generated the data, providing the other with only the mortality and TSP series to carry out the estimation. The parameter values were estimated correctly to within very small tolerances. The accuracy of estimates of the unobserved population is apparent in Figure 1, where the generated "actual" population is plotted along with the KF estimates of population over an arbitrary sequence of 400 days. The latter tracks both the local trend and day-to-day fluctuations. The KF requires a prior distribution for the unobserved state variable P_t at t = 0. The results were not sensitive to modifications of this prior distribution. We found that the estimated P₄ moved very quickly from its initial guess value to the correct level. These results give us added confidence in the validity of our results.

Results for the Philadelphia Mortality Data

We apply the model to daily data for Philadelphia from the period 1973–1990 provided to us by the Electric Power Research Institute. The data set includes daily observations on the number of non-accident deaths of individuals over age 65, a measure of TSP, and other atmospheric variables including temperature, barometric pressure, and ozone. Our investigation focused on the relationship of mortality to TSP. Before we were able to estimate the model, we had to address the problem of missing observations in the TSP series. Eliminating periods at the beginning and end of the sample where there are large gaps in the TSP record leaves a sequence of 5136 consecutive days within which there are only 50 days missing TSP values. To fill in these gaps,

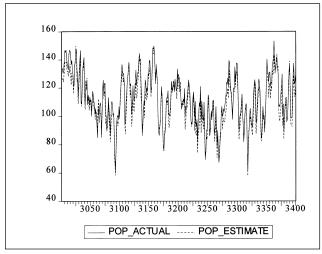


Figure 1. Comparison of actual at-risk population and KF estimate for simulated data

we regressed TSP on other atmospheric variables, using the integer part of the fitted series.

Table 1 displays maximum likelihood estimates of the observation and state equations for six baseline specifications of the x vector that include various combinations of TSP and AVGT, the square of temperature, and the multiplicative interaction of temperature with TSP. We chose AVGT as a baseline variable because of its strong linear association with daily mortality. A constant term is included in every case. Model 1 uses only TSP, which is highly significant. Model 2 adds AVGT, which is also highly significant. Model 3 adds the square of AVGT to allow for a hazard rate that increases at both extremes of temperature, which is also significant. Model 4 allows the effect of TSP and AVGT to each depend on the value of the other by adding the interaction variable AVGT*TSP. That interaction is significant, while TSP and AVGT alone are much less significant than in Model 3. Model 5 is included for comparison purposes and uses only the AVGT variables. Model 5 has roughly the same log likelihood as Model 1, which used only TSP. Finally, Model 6 uses all of the variables considered in this table: TSP and AVGT, the square of AVGT, and the interaction AVGT*TSP.

How important is TSP as a predictor of mortality? When comparing the log likelihood for Model 6 with that of Model 5, which does not include either TSP-related variable, the difference in value is about 14, which is highly significant (*p*-value is off the chi-square table). How important are the temperature-related variables? When comparing Model 6 with Model 1, the difference in log likelihoods is again about 14, and even with three additional parameters that difference is overwhelmingly significant. We also note that, although the interaction variable AVGT*TSP has a *t*-statistic of only 1.42 in Model 6, when we compare the log likelihood with that of Model 3, the difference is about 2. That corresponds to a *p*-value of about 0.05, supporting a role for the interaction term.

Thus, we regard Model 6 as a reasonable baseline specification. Note that to interpret the effect of TSP on the hazard rate in Model 6, one needs to keep in mind that the coefficient of TSP is a linear function of AVGT, so the effect is positive in spite of the negative coefficient of TSP alone. We also note that the residuals in the state equation lack serial correlation as assumed. For Model 6, those residuals display serial correlation of -0.012 at lag 1 day, -0.019 at lag 7 days, and 0.026 at lag 365 days, and none is statistically significant.

How do TSP and temperature affect the hazard rate? The hazard function of Model 6 implies that both extremes of temperature are harmful and that TSP is also harmful, with that effect rising with temperature. At a typical level of TSP, the effect of an increase in temperature from 40 °F to 90 °F is an increase in the hazard rate from 0.07 to 0.08. At a high temperature of 90 °F, the effect of an increase in TSP from zero to the highest levels observed in the sample is to raise the hazard rate from 0.075 to 0.091. Any statements about the effect of TSP on the hazard rate need to be tempered by the uncertainty connected with these estimates.

Figure 2 plots the KF estimate of the at-risk population and the estimated hazard rate in Model 6. The estimated at-risk population averages 484 and varies seasonally. Since average mortality is 35 per day, this implies that about 7% of the at-risk population dies on average per day. The corresponding time series of the hazard rate in Figure 2 is stable around its mean of 0.073, remaining well within the required (0,1) interval for a probability. The hazard rate fluctuates seasonally as periods of high TSP and temperature extremes reap a grim harvest, followed by less lethal conditions. As noted above, the mortality series is essentially trendless, so it is of some interest that the estimated at-risk population series does move higher with time, from an annual average of around 470 in early years to around 500 at the end. There is a

Table 1. Parameter estimates for baseline state-space models, Philadelphia data.

Coefficient	Model 1	Model 2	Model 3	Model 4	Model 5	Model 6
1	0.0632924 (8.61)	0.0650052 (9.34)	0.0696869 (8.95)	0.0686944 (7.72)	0.0765619 (8.94)	0.0711416 (1.89)
TSP	0.0000534 (5.60)	0.0000480 (5.50)	0.0000412 (4.09)	-0.000023 (-0.69)	n.a.	-0.000007 (-0.21)
AVGT	n.a.	0.0001002 (2.96)	-0.000167 (-2.95)	0.0000079 (0.62)	-0.000213 (-1.88)	-0.000168 (-0.21)
AVGT ²	n.a.	n.a.	0.0000030 (4.34)	n.a.	0.0000041 (3.35)	0.0000023 (0.28)
AVGT*TSP	n.a.	n.a.	n.a.	0.0000013 (2.28)	n.a.	0.0000009 (1.42)
σ	5.80991 (76.69)	5.756863 (74.90)	5.754371 (74.15)	5.755896 (73.14)	5.731345 (73.79)	5.754223 (49.13)
σ	19.7923 (13.63)	19.04966 (14.81)	19.05996 (13.95)	19.18479 (13.39)	18.58655 (14.78)	19.15221 (5.96)
N ["]	35.0947 (126.84)	35.08667 (131.80)	35.08826 (131.73)	35.08771 (130.85)	35.08178 (135.08)	35.08867 (131.02)
InL	-16924.9	-16919.03	-16913.12	-16914.00	-16924.72	-16910.97

Notes: t-statistics for the null hypothesis that an individual coefficient is zero are in parentheses beside it. The statistic InL is the log of the likelihood function.

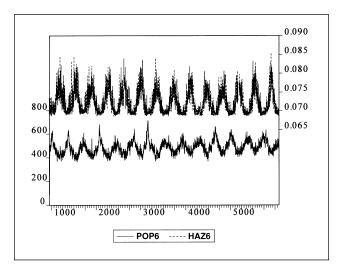


Figure 2. Estimated hazard rate and KF estimate of at-risk population in Model 6.

corresponding and offsetting decline in the hazard rate, moving downward from an annual average of about 0.072 in early years to 0.0716 in later years. This decline in the hazard rate is driven by TSP declining from annual averages of about 80 to about 70, the result of efforts to clean up the air around Philadelphia. A lower hazard rate lengthens life expectancy and allows individuals to stay longer in the at-risk population, thereby making that population larger. Finally, we note that output from all six models is qualitatively similar in producing population and hazard rate estimates that are stable and positive.

Does TSP have a cumulative effect on mortality over several days? We investigated this possibility by adding the level of TSP in preceding days to the list of variables in the hazard function in Models 1-3. The results for adding 4 or 8 lags of TSP are displayed in Table 2, which differs from Table 1 by adding another line listing the likelihood ratio statistic for testing the null hypothesis that all the added lagged values of TSP have a zero coefficient. What the test results indicate is that lags of TSP do have explanatory power that is statistically significant, but the coefficient estimates suggest that the effect is not large. The coefficient of current-day TSP is reduced when lags are added; the first lag is positive and significant, but the second lag is negative and remaining lags have small estimated coefficients. The sum of all TSP coefficients for Model 3 is raised, however, by almost half when 8 lags are added. How do we understand a negative effect of TSP after two days? One explanation is that the apparent effect of lags of TSP is partially statistical rather than biological. If we measure TSP imperfectly, due to aggregation over the 24-hr day or extrapolation of atmospheric measurements across a geographical region, then TSP on adjacent days may contain additional information relevant to the explanation of mortality. The lag pattern we obtain then may represent, at least in part, a kind of moving average measure of TSP rather than a biological process. Both factors may well be at work in producing the pattern seen here.

Table 2. Parameter estimates for state-space models with lags of TSP, Philadelphia data.

Coefficient	Model 1a	Model 1b	Model 2a	Model 2b	Model 3a	Model 3b
1	0.061616 (8.62)	0.057272 (8.30)	0.063138 (9.86)	0.058708 (8.76)	0.067901 (10.85)	0.0632121 (16.03)
TSP	0.000043 (4.64)	0.000042 (1.62)	0.000040 (4.76)	0.000040 (4.43)	0.000033 (4.31)	0.000033 (4.58)
TSP(-1)	0.000034 (4.11)	0.000033 (0.94)	0.000030 (3.73)	0.000030 (3.18)	0.000028 (3.33)	0.000027 (n.a.)
TSP(-2)	-0.000011 (n.a.)	-0.000011 (-0.22)	-0.000013 (n.a.)	-0.000013 (-4.51)	-0.000015 (-2.38)	-0.000014 (n.a.)
TSP(-3)	0.000009 (2.26)	0.000009 (0.51)	0.000009 (6.11)	0.000008 (n.a.)	0.000007 (n.a.)	0.000007 (n.a.)
TSP(-4)	0.000007 (3.20)	0.000004 (0.41)	0.000005 (n.a.)	0.000002 (0.38)	0.000004 (3.03)	0.000002 (n.a.)
TSP(-5)	n.a.	0.000005 (0.39)	n.a.	0.000004 (0.68)	n.a.	0.000004 (n.a.)
TSP(-6)	n.a.	0.000010 (1.26)	n.a.	0.000010 (1.58)	n.a.	0.000011 (n.a.)
TSP(-7)	n.a.	-0.000006 (-0.68)	n.a.	-0.000006 (-1.31)	n.a.	-0.000006 (n.a.)
TSP(-8)	n.a.	0.000000 (0.01)	n.a.	-0.000000 (0.05)	n.a.	0.000000 (n.a.)
AVGT	n.a.	n.a.	0.000070 (2.58)	0.000058 (1.22)	-0.000018 (0.00)	-0.000177 (2.44)
AVGT ²	n.a.	n.a.	n.a.	n.a.	0.000003 (3.06)	0.000003 (3.41)
$\sigma_{_{\it e}}$	5.80483 (76.22)	5.82535 (79.24)	5.73976 (77.88)	5.79522 (74.44)	5.76622 (78.30)	5.79049 (81.75)
σູ້	19.5818 (13.79)	20.0188 (13.54)	19.1180 (15.99)	19.5942 (14.15)	19.1418 (16.43)	19.5993 (15.91)
N"	35.0924 (128.18)	35.0979 (124.58)	35.0866 (131.30)	35.0924 (128.04)	35.0877 (131.23)	35.0935 (128.01)
InL	-16906.7	-16884.9	-16904.1	-16882.8	-16898.6	16877.2
LR	36.4***	80.0***	38.4***	72.46***	29.04***	67.54***

Notes: LR is the likelihood ratio statistic for testing the null hypothesis that the coefficients on all lags of TSP are zero, and in each case LR is significant at a p-value below 0.01. Individual t-statistics are not available in cases where the standard error could not be computed. See also the notes to Table 1.

We have also experimented with adding other atmospheric variables to baseline Model 6. Barometric pressure on the same day enters the model with a negative and significant coefficient, while the coefficients on TSP and AVGT change only slightly. The resulting estimated population and hazard series are essentially indistinguishable from those of Model 6.

DISCUSSION AND CONCLUSIONS

In thinking about the dynamics of the state-space model, we have found it useful to compare it to the more familiar linear regression or distributed lag model linking mortality to risk factors. The latter have the form

$$D_{t} = \alpha + \sum_{l=0}^{K} \beta_{k} x_{t-k} + \varepsilon_{t}$$
 (4)

The impact of $x_{t,k}$ on mortality is given by fixed coefficients in eq 4, while in the state-space model the impact depends on the unobserved state variable P. Because this functional relationship is multiplicative, the state-space model is highly non-linear. In particular, the impact will be smaller if recent mortality has been high, since the atrisk population will have been reduced. This harvesting effect is small in the models we estimated, but it is very persistent since the rate of replenishment of the population by new arrivals is slow. Thus, the relation of past atmospheric conditions to mortality is highly non-linear in the state-space model.

If we nevertheless run regressions of the form of eq 4 on the Philadelphia data set with TSP as the risk factor, there is a consistent pattern in the coefficients obtained. For example, with lags up to 7 days the lag zero coefficient is 0.013, but the sum over lags 0 through 7 days is -0.041. The character of the results is not sensitive to including more lags and other variables; the leading coefficient on TSP is positive but the sum $\sum \beta_k$ is small or negative. Thus, a positive association between mortality and TSP on the same day is more than offset by a sequence of negative coefficients, as predicted by the harvesting effect. The state-space model also predicts that the sum of coefficients will be small, since over the long run the level of mortality is only determined by the rate of arrivals, not by risk factors. Thus the state-space model proposed here helps explain why multiple regression produces results that are otherwise hard to explain, namely an apparent negative effect of TSP on mortality after a lag.

While an increase in risk factors cannot increase mortality in the long run, life expectancy is the inverse of the hazard rate, so hazard-causing agents will shorten it. The range of hazard rates observed over the sample period is roughly 0.07 to 0.085. Since life expectancy is the reciprocal of the hazard rate, the former corresponds to a life expectancy in the at-risk population of 14.3 days and the latter to 11.8 days, a difference of about 2.5 days on average for the roughly 500 at-risk individuals in Philadelphia.

In summary, we feel that the state-space approach represents a very promising avenue of research in modeling the effects of air pollution. Its primary advantage comes from explicit recognition of the role of population dynamics in mortality and from making use of the KF to estimate the population model directly. The results reported here are preliminary but suggest that the approach is practical and provides explanations of otherwise puzzling results from conventional analysis.

ACKNOWLEDGMENTS

This research was supported by the Electric Power Research Institute. We are grateful to Fred Lipfert and Ron Wyzga, who provided invaluable information, suggestions, and insights throughout this project. We also received helpful comments from Richard Burnett, Robert Engle, Levis Kochin, Jeremy Piger, Richard Startz, Mark Wohar, and anonymous referees. The views expressed, however, are those of the authors who are solely responsible for the content of this article.

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