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BiOBr photocatalyzed decarboxylation of glutamic acid: reaction rates, intermediates and mechanism<sup>†</sup>

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The degradation of glutamic acid by BiOBr under both UV and visible irradiation was investigated and compared with degradation by  $TiO_2/UV$ . Analysis of the reaction rates and the distribution of intermediates was used to show that both BiOBr systems, unlike the  $TiO_2$  system, catalyze direct substrate oxidation by valance band holes.

Visible light photocatalysis has attracted continuing attention because of potential applications in water purification.<sup>1,2</sup> A recently developed photocatalyst, BiOBr, is an efficient visible light photocatalyst<sup>3–5</sup> with visible light activity higher than that of N-doped TiO<sub>2</sub>.<sup>6</sup> As described previously, BiOBr has two discreet valance bands produced from O-2p and Br-4p orbitals.<sup>7,8</sup> The two bands respond, respectively, to UV and visible light excitation and holes of different oxidation potential are generated, providing multiple mechanisms for photocatalytic degradation.<sup>7</sup>

Amino acids are biologically important organic compounds, both as building blocks for proteins and as metabolic intermediates. As biodecomposition products, amino acids are distributed widely in natural waters.<sup>9,10</sup> The concentration of amino acids in surface water is generally in the range of 2.5–60 nM.<sup>11,12</sup> Although amino acids are nontoxic, they can form carcinogenic and mutagenic species during the water purification process.<sup>13,14</sup> For example, amino acids were converted primarily to halomethanes and haloacetic acids by chlorination.<sup>15–17</sup> More importantly, with naturally occurring toxins it is usually the carboxyl group of an amino acid that binds to the affected enzyme.<sup>18–21</sup> For example, with the wellknown cyanotoxin, microcystin-LR, the free carboxyl groups

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on D-Glu and D-MeAsp bind with the metal atom and Arg96 of protein phosphatase 1 (PP1) to inhibit protein phosphorylation.<sup>19,20</sup> Thus, understanding the degradation mechanism of amino acids, particularly the decarboxylation process, is of practical significance for water purification.

Glutamic acid (Glu) is one of the proteinogenic amino acids and, with a second carboxyl group on the side chain, it is an ideal substrate for comparing the degradation process of carboxylic acids with that of amino acids. In this work, we used BiOBr as the photocatalyst to degrade Glu under both UV and Vis irradiation. The degradation process was examined with <sup>1</sup>H NMR and <sup>18</sup>O isotope labeling and spin trapping ESR were used to elucidate the reaction mechanism. These results were compared with those from a TiO<sub>2</sub> system to show the effect of the valance band structure of BiOBr on the catalytic degradation of amino acids.

 $D_2O$  suspensions containing BiOBr and Glu were irradiated with UV or visible light for a given time and then analyzed using <sup>1</sup>H NMR analysis after removing the photocatalyst. Compared with parent substrate (Glu), the reacted solutions gave additional peaks at  $\delta$  of 1.93, 2.48, 3.26 and 8.35 with both UV and visible light irradiated systems (Fig. 1). Using reference



Fig. 1  $^{1}$ H NHR spectra of oxidative products of Glu in BiOBr/Vis and BiOBr/UV systems, 1 g L $^{-1}$  BiOBr,  $c_{Glu}^{0} = 10$  mmol L $^{-1}$ , 10 mL D<sub>2</sub>O.

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Fig. 2 Concentration change of substrate and oxidation products during photocatalytic oxidation of Glu in (a) BiOBr/Vis system and (b) BiOBr/UV system.

compounds, these peaks were assigned to acetic acid (AA), succinic acid (SA), malonic acid (MA) and formic acid (FA). The change in the <sup>1</sup>H NMR spectrum with reaction time also provides the kinetics of substrate consumption and intermediate formation in the BiOBr/UV and BiOBr/Vis systems (Fig. 2).

During photocatalytic oxidation, Glu forms SA initially and further reaction of the primary intermediate gives MA, AA and FA. However, no signal for aspartic acid, the decarboxylation product of Glu, was recorded, a clear indication that degradation begins with the amino group rather than the carboxyl. The different decarboxylation of Glu in BiOBr/Vis and BiOBr/UV systems was proposed in Scheme 1 (additional details are shown in Fig. S1 (ESI†)). Similarly, during the BiOBr photocatalyzed oxidation of microcystin-LR,<sup>22</sup> degradation also



Scheme 1 Difference in the decarboxylation of Glu in BiOBr/Vis (gray arrows) and BiOBr/UV (black arrows) systems. Solid arrows shows a major reaction route, dashed arrows represent a minor reaction route.

begins with oxidation of the amino-carboxyl structure of Glu. These results indicate that the amino-carboxyl structure is susceptible to oxidation in BiOBr photocatalytic systems. Due to the reactivity of this structure, amino acids are more readily degraded by BiOBr than are free carboxylic acids.

The BiOBr/Vis and BiOBr/UV systems display different degradation kinetics, but the visible light irradiated system also shows lower selectivity for SA and a markedly higher selectivity for MA (Table 1). This phenomenon is attributed to the discrete valance band structure of BiOBr. The holes generated by UV  $(h_{\text{O-2p}}^{+})$  and visible light  $(h_{\text{Br-4p}}^{+})$  excitation have different oxidation potentials, leading to different secondary reactions and the observed differences in rate. The results obtained in these systems were also compared with those of the classic TiO<sub>2</sub>/UV system to show the unique properties of BiOBr photocatalysis. It was observed that the TiO<sub>2</sub>/UV photocatalyzed oxidation of Glu gave remarkably low intermediate concentrations. The total selectivity of SA and MA in the TiO<sub>2</sub> photocatalyzed system is only 2.8%, which is much lower than that of BiOBr/Vis and BiOBr/UV systems (37.4% and 21.8%, respectively, Table 1). TiO<sub>2</sub> has a valance band ( $E_{\rm vb} = 2.7$  V) more oxidizing than either of the two valance bands of BiOBr and its hole oxidizes H<sub>2</sub>O to 'OH. It was reported that 'OH plays a significant role in TiO<sub>2</sub> photocatalyzed degradation of amino acids.23-25 Considering the valance band potentials and differences observed between the BiOBr and TiO<sub>2</sub> systems, we assume that the valance band hole of both BiOBr systems initiates the degradation of Glu by direct oxidation rather than by 'OH mediated reactions.

Since the photocatalytic degradation of Glu starts from the amino-carboxyl end and leads initially to SA, the decarboxylated and deaminated product, we anticipated that the source of oxygen atoms in the carboxyl group formed in this process could give useful information about the mechanism of the reaction. These experiments were carried out in <sup>18</sup>O-enriched water ( $H_2$ <sup>18</sup>O) and atmospheric <sup>16</sup>O<sub>2</sub>. Samples from the three systems were collected at times that resulted in similar substrate conversion (20–30%), and analyzed by derivative GC-MS (Fig. S2–S4 (ESI†)). As shown in Table 2, O atoms from both  $H_2O$  and  $O_2$  were incorporated into SA under BiOBr photocatalysis condition. The SA formed in BiOBr/UV and BiOBr/Vis systems have similar isotope abundances of carboxyl O atoms (<sup>16</sup>O% = 13–14), which illustrates that these two systems react with similar mechanisms. In contrast, the SA formed in

 
 Table 1
 The formation rate and selectivity of intermediates produced in the photocatalytic degradation of Glu

System	$r_{d}^{a} (\text{mmol } \text{L}^{-1} \text{ h}^{-1})$	$r_{f}^{b} (\mathrm{mmol}  \mathrm{L}^{-1}  \mathrm{h}^{-1})$		Sel. <sup>c</sup> (%)	
		SA	MA	SA	MA
BiOBr/Vis	0.199	0.019	0.055	9.6	27.8
BiOBr/UV	1.071	0.152	0.081	14.2	7.6
$TiO_2/UV$	13.33	0.197	0.173	1.5	1.3

<sup>*a*</sup> Decomposition rate of Glu. <sup>*b*</sup> Formation rate of intermediate. <sup>*c*</sup> Ratio of consumption rate of substrate to accumulation rate of intermediate.

Table 2Average isotope abundances of oxygen atoms in the carboxylgroup of SA in  $H_2^{18}O$  isotope labeling experiments<sup>a</sup>

System	Time (min)	Substrate conv. (%)	SA yield (%)	Abundance <sup>b</sup> (%)	
				<sup>16</sup> O <sub>2</sub>	H <sub>2</sub> <sup>18</sup> O
BiOBr/Vis	480	29.9	11.9	14.2	85.8
BiOBr/UV	90	23.8	23.8	13.1	86.9
$\mathrm{TiO}_{2}/\mathrm{UV}$	20	20.8	20.8	6.4	93.6

 $^a$  1 g L<sup>-1</sup> photocatalyst,  $c_{\rm Glu}^0 = 10 \text{ mmol L}^{-1}$ , 2 mL H<sub>2</sub><sup>18</sup>O.  $^b$  Average value of the two O atoms of the formed carboxyl group, corrected with the oxygen isotope abundance of solvent H<sub>2</sub><sup>18</sup>O and the natural isotope abundance of aerial O<sub>2</sub>.

the TiO<sub>2</sub>/UV system gave an <sup>16</sup>O abundance (<sup>16</sup>O<sub>9</sub> = 6.4) less than half that of the BiOBr systems. TiO<sub>2</sub> photocatalysis clearly incorporates more H<sub>2</sub>O derived oxygen to the product than the BiOBr systems. We also performed <sup>18</sup>O<sub>2</sub> isotope labeling experiments and similar results were obtained (Table S1 and Fig. S5–S7 (ESI†)). Since the valance band hole of TiO<sub>2</sub> can oxidize H<sub>2</sub>O to 'OH and incorporate O atoms from H<sub>2</sub>O to the product, the higher proportion of H<sub>2</sub>O derived oxygen in the TiO<sub>2</sub>/UV system is reasonable. These results also corroborate the direct oxidation mechanism proposed for BiOBr systems. We propose that, in both the BiOBr/UV and BiOBr/Vis systems,



**Fig. 3** ESR signals of the DMPO-<sup>•</sup>OH adducts in TiO<sub>2</sub>/UV, BiOBr/UV and BiOBr/Vis systems (a) without and (b) with Glu (10 mmol L<sup>-1</sup>). 1 g L<sup>-1</sup> photocatalyst,  $c_{_{DMPO}} = 0.4$  mol L<sup>-1</sup>.

the photogenerated hole  $(h_{O-2p}^+ \text{ or } h_{Br-4p}^+)$  oxidizes Glu to a cation radical, which then reacts with either  $O_2$  or  $H_2O$  to produce the carboxyl group on the product.

To further confirm that direct oxidation of Glu accounts for the larger pool of intermediates and higher proportion of O<sub>2</sub>derived oxygen in SA observed in the BiOBr systems, spintrapping ESR spectroscopy was used to detect the formation of 'OH. The results were again compared with those of TiO<sub>2</sub> and are shown in Fig. 3. In contrast to the TiO<sub>2</sub>/UV system, the signals from trapped 'OH recorded in the BiOBr systems was either weaker or nonexistent. Because neither of the valance band holes of BiOBr can oxidize H<sub>2</sub>O, the small amount of 'OH is attributed to the reduction of O<sub>2</sub> by conduction band electrons (O<sub>2</sub>  $\rightarrow$  'OOH  $\rightarrow$  H<sub>2</sub>O<sub>2</sub>  $\rightarrow$  'OH) and 'OOH was detected (Fig. S8 (ESI†)).

#### Conclusions

In summary, we studied the BiOBr catalyzed degradation of Glu under UV and visible light irradiation. Results indicate that, in both BiOBr/UV and BiOBr/Vis systems, the degradation process is initiated by direct substrate oxidation by the valance band hole. This, in turn, leads to the same primary product with the same source of oxygen in the carboxyl group formed on SA. However, the difference in the hole oxidation potentials of BiOBr/UV and BiOBr/Vis leads to different degradation rates, different secondary degradation processes and different distributions of degradation intermediates.

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#### Notes and references

- 1 R. Asahi, T. Morikawa, T. Ohwaki, K. Aoki and Y. Taga, *Science*, 2001, **293**, 269.
- 2 Z. Zou, J. Ye, K. Sayama and H. Arakawa, *Nature*, 2001, **414**, 625.
- 3 J. Li, Y. Yu and L. Zhang, Nanoscale, 2014, 6, 8473.
- 4 H. F. Cheng, B. B. Huang and Y. Dai, *Nanoscale*, 2014, 6, 2009.
- 5 S. L. Wang, W. H. Ma, Y. F. Fang, M. K. Jia and Y. P. Huang, *Appl. Catal.*, B, 2014, **150–151**, 380.
- 6 J. Wang, Y. Zhang, L. Tian, F. Liu and Q. Xia, *J. Nanopart. Res.*, 2014, **16**, 2691.
- 7 Y. F. Fang, W. H. Ma, Y. P. Huang and G. W. Cheng, *Chem. Eng. J.*, 2013, **19**, 3224.
- 8 S. L. Wang, L. L. Wang, W. H. Ma, D. M. Johnson, Y. F. Fang, M. K. Jia and Y. P. Huang, *Chem. Eng. J.*, 2015, **259**, 410.
- 9 G. Absalan, M. Akhond and L. Sheikhian, *Amino Acids*, 2010, **39**, 167.
- 10 J. T. Edward, P. G. Farrell and J. L. Job, *J. Am. Chem. Soc.*, 1974, **96**, 902.
- 11 K. Tada, M. Tada and Y. Maita, J. Oceanogr., 1998, 54, 313.
- 12 N. O. G. Jørgensen, Limnol. Oceanogr., 1987, 32, 97.

- 13 C. Goeschen, N. Wibowo and J. M. White, *Org. Biomol. Chem.*, 2011, **9**, 3380.
- 14 D. P. Li, C. Y. Hu, Y. L. Lin and S. J. Xia, *Sci. Total Environ.*, 2011, **409**, 1116.
- 15 T. Bond, J. Huang, M. R. Templeton and N. Graham, *Water Res.*, 2011, 45, 4321.
- 16 H. C. Hong, M. H. Wong and Y. Liang, Arch. Environ. Contam. Toxicol., 2009, 56, 638.
- 17 T. Bond, N. H. M. Kamal and T. Bonnisseau, J. Hazard. Mater., 2014, 278, 288.
- 18 K. J. Ullrich, G. Rumrich, Th. Wieland and W. Dekant, *Pflugers Arch., EJP*, 1989, **415**, 342.
- 19 A. Campos and V. Vasconcelos, *Int. J. Mol. Sci.*, 2010, **11**, 268–287.

- 20 C. Drahl, B. F. Cravattn and E. J. Sorensen, *Angew. Chem., Int. Ed.*, 2005, **44**, 5788.
- 21 A. Moshnikova, V. Moshnikova, O. A. Andreev and Y. K. Reshetnyak, *Biochemistry*, 2013, **52**, 1171.
- 22 Y. F. Fang, Y. P. Huang, J. Yang, P. Wang and G. W. Cheng, *Environ. Sci. Technol.*, 2011, **45**, 1593.
- 23 S. Horikoshi, N. Serpone, J. Zhao and H. Hidaka, J. Photochem. Photobiol., A, 1998, 118, 123.
- 24 L. Elsellami, F. Vocanson, F. Dappozze, R. Baudot, G. Febvay, M. Rey, A. Houas and C. Guillard, *Appl. Catal., B*, 2010, **94**, 192.
- 25 M. Matsushita, T. H. Tran, A. Y. Nosaka and Y. Nosaka, *Catal. Today*, 2007, **120**, 240.